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# LANDSCAPE FEATURES ARE A BETTER CORRELATE OF WILD PLANT POLLINATION THAN AGRICULTURAL PRACTICES IN AN INTENSIVE CROPPING SYSTEM

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1           LANDSCAPE FEATURES ARE A BETTER CORRELATE OF WILD PLANT POLLINATION THAN  
2                            AGRICULTURAL PRACTICES IN AN INTENSIVE CROPPING SYSTEM

3

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16

17   Abstract

18   Organic farming is commonly associated with increased pollinator diversity and abundance,

19   but the net effects on pollination rates are less documented. Besides, organic farms are often

20   surrounded by more diverse landscapes than conventional farms, such that the contributions

21   of landscape diversity vs. farming practices to pollination rates are often confounded with

22   each other. Here, the roles of local vs. landscape scale variables on pollination rates of

23   experimental plants are examined in agricultural landscapes. To this end, fruit set and seed

24   production were measured in the obligate insect-pollinated *Lotus corniculatus*. Plants were

25   located in pairs of neighbouring organic vs. conventional farms, which were characterized by

1 contrasting landscape structures and compositions. Fruit set, a proxy for pollinator visitation  
2 rates, was significantly related to landscape variables: fruit set was higher in farms close to a  
3 patch of semi-natural habitat, but lower in landscapes with a high cover of semi-natural  
4 habitats. Fruit set also correlated with local variables, such as habitat type, but not with  
5 farming type. Identical pollination rates in conventional and organic farms are likely due to  
6 similar diversities of habitats, crops and weeds in both farming types of the study area. These  
7 results therefore confirm that habitat diversity prevails over pesticide-free practices to explain  
8 the higher pollinator abundances usually observed in organic vs. conventional farms.

9

10 Keywords: *Lotus corniculatus* L.; fruit set; bumblebees; organic farming; semi-natural  
11 habitats

12

13 Highlights

- 14 • The effects of farming type vs. landscape on wild plant pollination were studied
- 15 • Pollination correlated with local habitat and presence of semi-natural habitats
- 16 • In contrast, pollination rates were similar in organic vs. conventional farms
- 17 • Habitat quality appears as the main factor enhancing pollination in organic farms

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19

## 1 1. Introduction

2           The ongoing loss of biodiversity in agricultural landscapes (e.g. Robinson and  
3 Sutherland, 2002) is believed to alter animal pollination, an ecosystem service essential for  
4 food production (Deguines et al., 2014; Klein et al., 2007) and for the sexual reproduction of  
5 many wild plants (Kearns et al., 1998; Ollerton et al., 2011). Beyond the simple effect of  
6 reduced numbers of pollinators, the pollination crisis may also be related to declines in the  
7 functional diversity of pollination networks, which can lower reproductive success and  
8 community persistence (Fontaine et al., 2006). Pollinator declines have been reported  
9 numerous times (Potts et al., 2010 for reviews; Steffan-Dewenter et al., 2005; Winfree et al.,  
10 2009), particularly in intensive agricultural landscapes, and the expected parallel declines of  
11 insect-pollinated plants are already observed at large scales in Europe (Biesmeijer et al.,  
12 2006). Plants provide food and habitats for many animal species involved in biological  
13 control for example, so that pollinator-induced changes in plant communities and their  
14 diversity could also have cascading effects on other ecosystem services (Scherber et al.,  
15 2010).

16           The design of efficient conservation schemes to reverse the loss of pollinators and  
17 pollination services in agroecosystems requires a complete understanding of the mechanisms  
18 responsible for this downward trend. Positive effects of organic farming on pollinator species  
19 richness and abundance have been documented numerous times (Clough et al., 2007;  
20 Holzschuh et al., 2010, 2008, 2007; Kennedy et al., 2013; Kremen et al., 2002; Rundlöf et al.,  
21 2008a, 2008b; Rundlöf and Smith, 2006), which suggests that agricultural practices in  
22 conventional fields are partly responsible for pollinator loss. Pesticide-free practices, together  
23 with abundant and diverse floral resources, likely provide higher quality crop and non-crop  
24 habitats and food sources in organic vs. conventional farms. Organic farms could therefore

1 sometimes sustain populations of wild pollinators without requiring semi-natural habitats,  
2 thanks to high-quality foraging and nesting sites (Williams and Kremen, 2007).

3         However, the effective impact of organic farming on pollinators and pollination is still  
4 open to discussion. First, the actual effects of organic farming often depend on the landscape  
5 context, with higher benefits in intensive landscapes (Holzschuh et al., 2007; Rundlöf and  
6 Smith, 2006). Second, organic farms often encompass a larger proportion of semi-natural  
7 habitats (e.g. grassland, field boundaries or hedgerows) than conventional farms (Feber et al.,  
8 2007; Fuller et al., 2005; Gibson et al., 2007; Norton et al., 2009). Pollinator abundance and  
9 diversity are known to depend on the presence and proximity of semi-natural habitats  
10 (reviewed in Ricketts et al., 2008), which provide both nesting sites and foraging resources, so  
11 that the observed positive effects of organic farming on pollinators could be mostly  
12 attributable to a landscape effect, which has considerable implications to design conservation  
13 schemes. Finally, most studies of the impact of agricultural practices have focused on  
14 pollinator diversity and abundance (Andersson et al., 2012; but see e.g. Brittain et al., 2010b;  
15 Carvalheiro et al., 2010), whereas pollinator abundance and pollination efficiency are  
16 sometimes only weakly correlated (Ricketts et al., 2008). There is thus a need to (1) further  
17 disentangle the relative effects of in-field agricultural practices (influencing the quality of the  
18 agricultural matrix as a foraging and nesting resource) vs. landscape features (isolation from  
19 semi-natural habitat) on pollination, and (2) measure actual pollination efficiency as a  
20 necessary complement to pollinator abundance.

21         Here, the pollination rate of potted birdsfoot trefoil plants (*Lotus corniculatus* L.,  
22 Fabaceae), estimated in terms of fruit and seed set, was compared across contrasting  
23 landscapes using a paired design (eight pairs of organic/conventional farms) in an intensive  
24 agricultural region in France. The following questions were specifically addressed: (1) Is wild  
25 plant pollination higher in organic vs. conventional farms regardless of landscape features? If

1 organic practices favour pollinator survival and reproduction in the agricultural matrix, a  
2 positive effect of organic farming is expected irrespective of the distance to or proportion of  
3 semi-natural habitats in the surroundings. (2) How do semi-natural habitats influence  
4 pollination rates in farms? An increase in pollination rates is expected closer to semi-natural  
5 habitats (if the latter act as sources of pollinators) or with larger proportions of semi-natural  
6 habitats (as providers of food resources), at least in conventional farms.

7

## 8 2. Material and methods

9 The pollination of experimental plants was monitored in eight pairs of crop farms in the  
10 intensive agricultural region surrounding Paris, France. Each pair consisted of one  
11 conventional and one organic farm located close to each other (supplementary Fig. 1, mean  
12 distance between paired farms 2915 m; range: 490-6000 m). A farm is defined as a collection  
13 of fields managed by the same person, hence under identical practices. Organic farms are rare  
14 in the study region, such that their fields are often interspersed with conventional fields. To  
15 avoid confounding effects of neighbouring agricultural practices (e.g. contamination by  
16 pesticides from an adjacent conventional field) eight organic farms consisting of clustered  
17 fields were sampled first. A conventional farm of similar size was then selected within five  
18 km of each organic farm (supplementary Fig. 1). The eight pairs of farms were located in  
19 contrasting landscapes, which were described a posteriori by the quantity and proximity of  
20 semi-natural habitats (see section 2.2).

21 All farms grew mainly cereals, the dominant crops in the study region, except one  
22 organic farm that also included a few grasslands and a small sheep herd. Organic farms grew  
23 a significantly larger number of crops than conventional farms (Supplementary Table 1; mean  
24  $\pm$  SD:  $9 \pm 2.4$  in organic farms vs.  $5.6 \pm 1.1$  in conventional farms, Student's test,  $P = 0.005$ ).  
25 Some of the supernumerary crops grown in organic farms are pollinator-attractive plants, so

1 that one may expect higher pollinator visitation rates in organic farms. Note however that  
2 pollinator-attractive crops represented a small proportion of arable land in both farming  
3 system, and that conventional farms also grew pollinator-attractive crops (e.g. oilseed rape,  
4 Supplementary Table 1).

5 In each farm two experimental sites were selected in each of two habitats: set-asides  
6 and margins of cereal fields. One conventional farm (C6) contained no set-aside that year; all  
7 sites were thus located in margins of cereal fields. Set-asides contain numerous pollinator-  
8 attractive flowers in contrast to margins of cereal fields; the two habitats can be compared to  
9 evaluate the influence of local resource availability on pollination rates. Within a farm,  
10 experimental sites were chosen so that they were widely distributed over the farm area, yet  
11 never adjacent to a conventional field in organic farms (and vice versa) or to mass-flowering  
12 crops such as oilseed rape or buckwheat. No managed honeybee hive was present within the  
13 studied farms. The mean distance between two sites within a farm was 750 m (range: 175-  
14 2040 m).

15

### 16 2.1. *Measurement of pollination efficiency*

17 The self-compatible, strictly entomogamous (Ollerton and Lack, 1998) birdsfoot trefoil  
18 was used to quantify pollination by several *Bombus* species, which are abundant in the study  
19 area. The cultivated forage crop varieties Leo and Baco were chosen to minimize individual  
20 variation in morphological traits such as number of flowers per inflorescence. Wild *L.*  
21 *corniculatus* is common in the study region (although it was not found in the direct vicinity of  
22 the experimental sites, see below) but was not cultivated as a crop in the study area.

23 All plants were first grown together in an insect-proof greenhouse from April to June  
24 2009. A total of 256 pots, each containing three to five individuals of either variety (Leo, 128  
25 pots; Baco; 128 pots) were used for this experiment. Just before flowering, potted plants were

1 moved to experimental sites, with four individual pots per site (two pots of each variety to  
2 prevent potential self-incompatibility among individuals of the same variety). No native *L.*  
3 *corniculatus* plants were found in the close vicinity of the sites. After two weeks in the field,  
4 i.e. just before farmers started harvesting fields and mowing set-asides, plants were brought  
5 back to the insect-free greenhouse. Every inflorescence was labelled and the number of  
6 flowers counted. At pod maturity, fruits of each inflorescence were counted, harvested and  
7 opened to count seeds.

8           A substantial number of plants were lost in the field, due either to water deprivation  
9 or to the destruction of pots by farmers or wild boars: only 93 pots from 51 (instead of 64)  
10 sites were retrieved. This loss created two imbalances in the dataset: there were many more  
11 pots of the Baco than of the Leo variety (78 vs. 15) and data were available for one habitat  
12 only in two farms (C5 and C8, see Supplementary Fig. 1), in addition to farm C6. However,  
13 discarding the Leo variety or the farm pairs 5, 6 and 8 did not modify our main results.  
14 Results presented here are therefore based on the full dataset.

15           During the experiment, positive and negative controls were grown outside and inside  
16 the greenhouse. The negative controls consisted of 12 pots of each variety kept in the insect-  
17 free greenhouse during the flowering period. Inflorescences were labelled and flowers  
18 counted (total: 694 flowers on 200 inflorescences). At the end of the experiment, no fruit was  
19 found on any of the marked branches. The positive controls consisted of three sites of four  
20 pots (two of each variety, as in the farms) located in a grassland nearby the greenhouse,  
21 surrounded by natural habitats. In the positive controls, the number of fruits and seeds per  
22 fruit was counted on each inflorescence. The between-variety difference in fruit and seed  
23 production was tested using generalized linear mixed-effects models (glmer, R package  
24 lme4), with the number of fruits or seeds as response variables and with a Poisson error and



1 log-link function. The models included ‘Site’, ‘Pot’ (and ‘Inflorescence’ for seed production)  
2 as random factors, and ‘Variety’ as a fixed factor.

3

#### 4 *2.2. Landscape metrics*

5 We used digital land-use maps (Mode d’Occupation des Sols, Institute for Planning and  
6 Development of the Île-de-France Region 2003, <http://www.iaurif.org>) and ArcGIS 9.2  
7 software (ESRI 2006, Redlands) to identify artificial (arable land, other non-grassland  
8 agricultural land, built area and open urban area) vs. semi-natural habitats (woodland,  
9 perennial grassland, water bodies and non-agricultural rural land). The distance to the nearest  
10 patch of semi-natural habitat (range: 1-698 m) and the proportion of semi-natural habitat in  
11 circular buffers around each study point (250 m, 500 m, 750 m and 1000 m radius, range: 0-  
12 56 % of semi-natural habitat) were calculated for each experimental site with the Patch  
13 Analyst extension. Within farm pairs, there was no significant difference in landscape features  
14 (distance to semi-natural habitat and proportion of semi-natural habitats) between organic and  
15 conventional farms (linear models with ‘Farming type’ and ‘Habitat’ as fixed factors,  $P >$   
16 0.2).

17

#### 18 *2.3. Statistical analyses*

19 We first selected the most relevant landscape scale to study the relationship between  
20 pollination and proportion of semi-natural habitats around the sites following the approach of  
21 Holzschuh et al. (2008). To this end, single-factor generalized linear mixed-effect models  
22 (glmer, R package lme4) were used to model two response variables: (1) fruit set, the  
23 proportion of flowers setting fruits on an inflorescence, with a binomial error and a logit-link  
24 function and (2) seed set, the number of seeds per fruit, with a Poisson error and log-link  
25 function. Both types of models contained ‘Pot’, ‘Site’ and ‘Farm pair’ as nested random

1 factors; an ‘Inflorescence’ random factor was also included in the model for seed set. In this  
2 preliminary approach, the only fixed factor was ‘Proportion of semi-natural habitats’. The  
3 four buffer radii (250, 500, 750 and 1000 m) were analysed separately. The scale maximizing  
4 the variance calculated from the fixed effect component was retained: 250 m for fruit set and  
5 500 m for seed set.

6 The relationship between pollination and environmental variables was examined using  
7 the generalized linear mixed-effects models above, with the same random factors but more  
8 fixed factors. The models included the following main fixed effects: Farming system (organic  
9 vs. conventional), Local habitat (field margin vs. set-aside), Distance to the nearest semi-  
10 natural habitat, and Proportion of semi-natural habitats. A number of fixed-effect covariates  
11 potentially influencing pollination or seed production were added: *L. corniculatus* variety  
12 (Baco vs. Leo), Number of flowers per pot at harvesting (as a measure of pollinator attraction  
13 of a site), and Number of *L. corniculatus* varieties per site (which was 1 when both pots of a  
14 variety were destroyed in a site). The latter variable was used as a conservative estimate of  
15 availability of compatible pollen. For fruit set, the model also included an interaction between  
16 flower number and variety number. Two-way interactions between farming system and  
17 landscape metrics, as well as between farming system and local habitat, were never  
18 significant and were omitted without affecting the outcomes of the models. For both models  
19 marginal and conditional  $R^2$ , the percentage of variance explained by fixed factors alone or by  
20 fixed and random factors, respectively, were calculated using the formulae of Nakagawa and  
21 Schielzeth (2013) for generalized linear mixed-effects models.

22

### 23 3. Results

24 The negative controls grown in an insect-proof greenhouse confirmed that seed  
25 production in *L. corniculatus* requires visitation by pollinators: none of the 694 flowers

1 examined set fruit. In the positive controls with open pollination, the number of fruits per  
2 inflorescence and per pot differed between varieties (mean  $\pm$  SD:  $2.9 \pm 1$  fruits per  
3 inflorescence in Leo vs.  $1.9 \pm 1$  in Baco;  $20 \pm 19$  fruits per pot in Leo vs.  $11 \pm 8$  in Baco), but  
4 the number of seeds per fruit was similar ( $8.3 \pm 4$  in Leo vs.  $8.6 \pm 4$  in Baco).

5       Plants grown two weeks in field sites exhibited smaller fruit sets ( $0.7 \pm 1$  fruits per  
6 inflorescence;  $3.4 \pm 6$  fruits per pot) than positive controls, with no difference between  
7 varieties (Table 1). Fruit set correlated with some local field site characteristics: it was  
8 significantly higher in set-asides than in field margins and in sites where both *L. corniculatus*  
9 varieties remained until the end of the experiment than in sites where the two pots of a given  
10 variety were destroyed (Table 1). Overall, fruit set was not related with the total number of *L.*  
11 *corniculatus* flowers per sites, but there was a significant interaction between number of  
12 flowers and number of varieties (Table 1): fruit set was positively related to the number of  
13 flowers in sites with a single variety, but was unrelated to flower number in sites with two  
14 varieties. There was no difference in fruit set between organic and conventional farming,  
15 regardless of local habitat type and surrounding landscape (no significant interactions  
16 between farming system and local habitat type or landscape variables). Finally, fruit set was  
17 partly predicted by landscape variables (Table 1): it decreased with increasing the proportion  
18 of semi-natural habitats surrounding the sites and with increasing isolation from semi-natural  
19 habitats (Fig. 1).

20       In the fields, the number of seeds per fruit was comparable to that of the positive  
21 controls ( $8.5 \pm 6$  seeds per fruit) and was explained by none of the local or landscape-scale  
22 fixed factors (Table 1). Most variation occurred at a very fine scale, i.e. among inflorescences  
23 within a pot (conditional  $R^2$ ,  $R^2_{\text{GLMM}(c)} \approx 67\%$ , most of which is explained by variance among  
24 inflorescences, Table 1). Note that this fine scale variation also explained an appreciable  
25 fraction of variance in fruit set (large among-pots variance, Table 1).

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#### 4. Discussion

This study aimed to separate the relative effects of local farming practices (organic vs. conventional farming; field margin vs. set-aside habitat) and landscape (distance to and proportion of semi-natural habitat) on pollination provided to wild plants in agricultural areas. The comparison of pairs of farms located in the same landscape showed a combination of local and landscape-scale effects, but no detectable effect of organic vs. conventional farming, on the fruit set of *L. corniculatus*. The number of seeds per fruit was unrelated to any environmental variable.

In our experiment, fruit and seed set were used as proxys for pollination efficiency. Although no data on pollinator abundance and visitation rates were available (in contrast to Brittain et al., 2010b), part of variation in fruit set can be interpreted in terms of pollinator visitation rates. Bearing in mind that pollinator visitation is required for fertilization in *L. corniculatus* (no self-fertilization, as confirmed by the caged control plants setting no fruits, see also Ollerton and Lack, 1998) and that no wild conspecifics were found in the study area, differences in fruit set should only be caused by one of the following: (1) differences in pollinator visitation rates; (2) differences in heterospecific pollen loads on pollinators, which, when large, may dilute small amounts of *L. corniculatus* pollen and impede pollen transfer (Ashman and Arceo-Gomez, 2013); and (3) differences in the availability of compatible *L. corniculatus* pollen within sites. We observed the latter effect, i.e. a negative relationship between fruit set and compatible pollen availability, as estimated conservatively by the number of *L. corniculatus* varieties remaining in a site (Table 1). However, the number of varieties per site was unrelated to other environmental variables and should not be responsible for the relationship of the latter with fruit set. Below we therefore discuss how environmental

1 variables influence pollinator visitation and, to a lesser extent, pollen loads across spatial  
2 scales.

3

#### 4 4.1. *Local and landscape scale effects on fruit set*

5       Apart from the effect of the number of varieties per site, variation in fruit set was mostly  
6 explained by local variables related to availability of floral resources and by landscape  
7 variables; both effects can be interpreted in terms of pollinator attraction and abundance. At  
8 the local scale, fruit set was higher in the more flower-rich set-asides than in field margins  
9 and in sites with more *L. corniculatus* flowers (although the latter relationship was significant  
10 only in sites with a single *L. corniculatus* variety), regardless of farming system. In other  
11 words, sites with more floral resources (either *L. corniculatus* or other weed species) were  
12 more attractive to pollinators and received more visits, which is consistent with the well-  
13 known positive effects of floral resource availability on pollinator visitation rate and  
14 behaviour (Kunin, 1997).

15       At the landscape scale, fruit set decreased with increasing the distance to semi-natural  
16 habitats, which again is a well-documented pattern (see Ricketts et al., 2008 for a meta-  
17 analysis) that may be explained by higher visitation rates in the vicinity of semi-natural  
18 habitats. Pollinator abundance and richness are known to depend on the proximity of semi-  
19 natural habitats, which provide nesting sites for many pollinator species (Farwig et al., 2009;  
20 Ockinger and Smith, 2007), although the consequences on seed set are often weaker (Farwig  
21 et al., 2009; Kohler et al., 2008; Steffan-Dewenter and Tscharntke, 1999), especially with  
22 flowering crop pollination (Ricketts et al., 2008).

23       More surprisingly, the proportion of semi-natural habitats and fruit set were negatively  
24 related, which contradicted the general expectation that semi-natural habitats, as a source of  
25 pollinators, would increase pollination service (Kremen et al., 2004). This counterintuitive

1 pattern was however already observed elsewhere (Brittain et al., 2010b; Winfree et al., 2007)  
2 and could be attributable to several phenomena. The explanation put forward by Winfree et al.  
3 (2007), that semi-natural habitats may be of lower quality than agricultural habitats in terms  
4 of pollinator nesting sites and food sources, can immediately be dismissed here. Winfree et al.  
5 (2007) observed lower bee abundance in forest vs. agricultural or urban areas, which created a  
6 negative relationship between pollinator diversity or abundance and forest cover in the  
7 surrounding landscape. In the present study, pollination was generally higher next to semi-  
8 natural habitats (Fig. 1), which consisted in a combination of forests and grassland; these  
9 habitats are therefore unlikely to be of lower quality than intensively managed farmland.

10         Alternatively, semi-natural habitats may offer abundant foraging resources to  
11 pollinators; they would therefore be more attractive than agricultural areas. Small patches of  
12 plants isolated in an otherwise hostile agricultural matrix should therefore receive few visits if  
13 abundant resources are available in the surrounding semi-natural habitats (Dauber et al., 2010;  
14 Wilcock and Neiland, 2002). The attractive effects of semi-natural habitats should depend  
15 strongly on pollinator foraging distance (Wilcock and Neiland, 2002). Bumblebees, the main  
16 pollinators of *L. corniculatus*, are large insects with potentially long foraging distances, but  
17 they depend on the presence of nearby foraging resources (optimal foraging theory). Their  
18 foraging distance can thus become short (less than 300 m, the buffer radius in our study) if  
19 floral resources are available in the nest vicinity (Wolf and Moritz, 2008), so that they may  
20 prefer foraging in nearby flower-rich semi-natural habitats than in the flower-poor agricultural  
21 matrix. In addition, the diversity of floral resources associated with more abundant semi-  
22 natural habitats may increase pollen limitation via the deposition of heterospecific pollen  
23 (Wilcock and Neiland, 2002).

24

25 4.2. *Organic farming is not associated with increased fruit set*

1           The absence of a relationship between fruit or seed production in *L. corniculatus* and  
2 farming system may appear at odds with numerous previous observations of higher pollinator  
3 richness and abundance, and sometimes seed or fruit production, in organic vs. conventional  
4 farms (Andersson et al., 2012; Bengtsson et al., 2005 for a meta-analysis; Holzschuh et al.,  
5 2008; Kennedy et al., 2013). However, in these studies, organic farms sometimes differed in  
6 many respects from conventional farms, including field management (use of synthetic inputs),  
7 farm-scale diversity (number of crops, presence of non-crop vegetation) and landscape  
8 configuration or composition, all of which can influence pollinator abundance and behaviour.  
9 Importantly, these studies showed that local and landscape factors can interact, such that for  
10 example the benefits of organic management should be highest in simplified landscapes (e.g.  
11 Bengtsson et al. 2005; Kennedy et al. 2013) or landscape effects should be highest as farms  
12 become increasingly simplified (Kennedy et al. 2013). The former interaction was not  
13 observed in the present study (no significant interaction between landscape variables and  
14 farming type). However, this and other studies are consistent with a predominant role of  
15 landscape variables over farming system in simplified farms, characterized by low habitat or  
16 weed diversity: Kremen et al. (2002) in Californian watermelon fields, Winfree et al. (2008)  
17 in four crops of North-Eastern USA, Brittain et al. (2010b) in Italian vine fields, or  
18 Carvalheiro et al. (2010) in South-African mango fields. In this study, farming was intensive  
19 and characterized by a low diversity of crop and non-crop habitats, even in organic farms.  
20 Although the number of crops was higher in organic than in conventional farms, cereals were  
21 the dominant crop and pollinator-attractive crops represented small proportions of cropland in  
22 both farming types. Hedgerows, field margins, and permanent set-asides were rare and weeds  
23 were controlled in both farming systems, although with different methods (e.g. set-asides  
24 were frequently mown in organic farms). As a consequence, pollinator visits were generally  
25 rare in agricultural habitats, especially in margins of cereal field: fruit set, but not seed

1 production per fruit, was lower than that of positive controls, and lower than that of similar *L.*  
2 *corniculatus* pots located in urban areas (Pellissier et al., 2012).

3         This study therefore corroborates earlier results indicating that at a local scale, the  
4 beneficial effects of organic farming on pollinators are attributable to the diversity of habitat  
5 and food sources they provide, much more than to the absence of pesticides. For example,  
6 Holzschuh et al. (2008) showed that “species richness of flowering plants” or “flower cover”  
7 could be substituted for “farming system” in models explaining variation in pollinator  
8 abundance in fallow strips and detected no significant effect of insecticides. Similarly,  
9 Carvalheiro et al. (2011) demonstrated that the presence of weeds allowed pollinators to  
10 persist in sunflower fields, sometimes reversing the negative effects of distance to natural  
11 habitats. In the organic farms studied here, which were little diversified, exclusion of  
12 pesticides is therefore unlikely to benefit pollinators, because the latter do not use organic  
13 crops or non-crop elements much as habitats. Conversely, pesticide applications in  
14 conventional farms are likely to have limited, transient negative impacts on pollinators by  
15 affecting only a small fraction of foraging individuals (see Brittain et al., 2010a for an  
16 example of such transient decrease in pollinator visitation rates after pesticide application).  
17 Regardless of pesticide effects, the patterns observed here are consistent with the prediction of  
18 Kennedy et al. (2013) that transition to organic monocultures, with little on-farm habitat  
19 diversity, may be detrimental for pollinators.

20

### 21 *4.3. Limitations and perspectives*

22         Our study showing significant effect of landscape variables, but no effect of farming  
23 system, on pollination rates was restricted in space, time and pollination type. First, the  
24 dominant effect of landscape and semi-natural habitats on pollination rates may be of special  
25 significance for *L. corniculatus*, a specialist plant pollinated by a large generalist pollinator,



1 because many other plant species may satisfy the foraging needs of bumblebees in semi-  
2 natural habitats. Contrasting patterns could be observed for plants visited by different  
3 pollinator communities (e.g. generalist plants pollinated, among others, by specialist  
4 pollinators). For example, the effects of local practices may be stronger for smaller  
5 pollinators, which usually have smaller foraging distances (Benjamin et al., 2014). Second,  
6 there was no significant effect of organic farming on pollination over a two-week period,  
7 which suggests identical pollinator visitation rates over this time frame. However, pollination  
8 rates may differ at other times (e.g. just after pesticide applications, Brittain et al. 2010a),  
9 which could result in a lower cumulative reproductive success of wild plants over the longer  
10 term. Gabriel and Tschardt (2007) showed for example a higher proportion of insect-  
11 pollinated arable weeds in field edge and field centre of organic vs. conventional farms,  
12 suggesting such a long-term benefit of organic farming. This effect of organic farming could  
13 be mediated by the diversity of floral resources and habitats, but longer-term studies are  
14 nonetheless needed to assess plant reproductive success in both farming systems, with  
15 quantification of pollinator abundance, visitation rate and floral resources in cultivated as well  
16 as non-cultivated areas.

17

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## 1 **References**

- 2 Andersson, G.K.S., Rundlöf, M., Smith, H.G., 2012. Organic farming improves pollination  
3 success in strawberries. *Plos One* 7, e31599. doi:10.1371/journal.pone.0031599
- 4 Ashman, T.-L., Arceo-Gomez, G., 2013. Toward a predictive understanding of the fitness  
5 costs of heterospecific pollen receipt and its importance in co-flowering  
6 communities. *Am. J. Bot.* 100, 1061–1070. doi:10.3732/ajb.1200496
- 7 Bengtsson, J., Ahnstrom, J., Weibull, A.C., 2005. The effects of organic agriculture on  
8 biodiversity and abundance: a meta-analysis. *J. Appl. Ecol.* 42, 261–269.  
9 doi:10.1111/j.1365-2664.2005.01005.x
- 10 Benjamin, F.E., Reilly, J.R., Winfree, R., 2014. Pollinator body size mediates the scale at  
11 which land use drives crop pollination services. *J. Appl. Ecol.* 51, 440–449.  
12 doi:10.1111/1365-2664.12198
- 13 Biesmeijer, J.C., Roberts, S.P.M., Reemer, M., Ohlemueller, R., Edwards, M., Peeters, T.,  
14 Schaffers, A.P., Potts, S.G., Kleukers, R., Thomas, C.D., Settele, J., Kunin, W.E., 2006.  
15 Parallel declines in pollinators and insect-pollinated plants in Britain and the  
16 Netherlands. *Science* 313, 351–354. doi:10.1126/science.1127863
- 17 Brittain, C., Bommarco, R., Vighi, M., Barmaz, S., Settele, J., Potts, S.G., 2010a. The impact  
18 of an insecticide on insect flower visitation and pollination in an agricultural  
19 landscape. *Agric. For. Entomol.* 12, 259–266. doi:10.1111/j.1461-  
20 9563.2010.00485.x
- 21 Brittain, C., Bommarco, R., Vighi, M., Settele, J., Potts, S.G., 2010b. Organic farming in  
22 isolated landscapes does not benefit flower-visiting insects and pollination. *Biol.*  
23 *Conserv.* 143, 1860–1867. doi:10.1016/j.biocon.2010.04.029
- 24 Carvalho, L.G., Seymour, C.L., Veldtman, R., Nicolson, S.W., 2010. Pollination services  
25 decline with distance from natural habitat even in biodiversity-rich areas. *J. Appl.*

1 Ecol. 47, 810–820. doi:10.1111/j.1365-2664.2010.01829.x

2 Carvalho, L.G., Veldtman, R., Shenkute, A.G., Tesfay, G.B., Pirk, C.W.W., Donaldson, J.S.,  
3 Nicolson, S.W., 2011. Natural and within-farmland biodiversity enhances crop  
4 productivity. *Ecol. Lett.* 14, 251–259. doi:10.1111/j.1461-0248.2010.01579.x

5 Clough, Y., Holzschuh, A., Gabriel, D., Purtauf, T., Kleijn, D., Kruess, A., Steffan-Dewenter,  
6 I., Tschardtke, T., 2007. Alpha and beta diversity of arthropods and plants in  
7 organically and conventionally managed wheat fields. *J. Appl. Ecol.* 44, 804–812.  
8 doi:10.1111/j.1365-2664.2007.01294.x

9 Dauber, J., Biesmeijer, J.C., Gabriel, D., Kunin, W.E., Lamborn, E., Meyer, B., Nielsen, A.,  
10 Potts, S.G., Roberts, S.P.M., Sober, V., Settele, J., Steffan-Dewenter, I., Stout, J.C.,  
11 Teder, T., Tscheulin, T., Vivarelli, D., Petanidou, T., 2010. Effects of patch size and  
12 density on flower visitation and seed set of wild plants: a pan-European  
13 approach. *J. Ecol.* 98, 188–196. doi:10.1111/j.1365-2745.2009.01590.x

14 Deguines, N., Jono, C., Baude, M., Henry, M., Julliard, R., Fontaine, C., 2014. Large-scale  
15 trade-off between agricultural intensification and crop pollination services.  
16 *Front. Ecol. Environ.* 12, 212–217.

17 Farwig, N., Bailey, D., Bochud, E., Herrmann, J.D., Kindler, E., Reusser, N., Schuepp, C.,  
18 Schmidt-Entling, M.H., 2009. Isolation from forest reduces pollination, seed  
19 predation and insect scavenging in Swiss farmland. *Landsc. Ecol.* 24, 919–927.  
20 doi:10.1007/s10980-009-9376-2

21 Feber, R.E., Johnson, P.J., Firbank, L.G., Hopkins, A., Macdonald, D.W., 2007. A comparison  
22 of butterfly populations on organically and conventionally managed farmland. *J.*  
23 *Zool.* 273, 30–39. doi:10.1111/j.1469-7998.2007.00296.x

24 Fontaine, C., Dajoz, I., Meriguet, J., Loreau, M., 2006. Functional diversity of plant-  
25 pollinator interaction webs enhances the persistence of plant communities. *PLoS*

1 Biol. 4, 129–135. doi:10.1371/journal.pbio.0040001

2 Fuller, R.J., Norton, L.R., Feber, R.E., Johnson, P.J., Chamberlain, D.E., Joys, A.C., Mathews,  
3 F., Stuart, R.C., Townsend, M.C., Manley, W.J., Wolfe, M.S., Macdonald, D.W.,  
4 Firbank, L.G., 2005. Benefits of organic farming to biodiversity vary among taxa.  
5 Biol. Lett. 1, 431–434. doi:10.1098/rsbl.2005.0357

6 Gabriel, D., Tschardtke, T., 2007. Insect pollinated plants benefit from organic farming.  
7 Agric. Ecosyst. Environ. 118, 43–48. doi:10.1016/j.agee.2006.04.005

8 Gibson, R.H., Pearce, S., Morris, R.J., Symondson, W.O.C., Memmott, J., 2007. Plant  
9 diversity and land use under organic and conventional agriculture: a whole-farm  
10 approach. J. Appl. Ecol. 44, 792–803. doi:10.1111/j.1365-2664.2007.01292.x

11 Holzschuh, A., Steffan-Dewenter, I., Kleijn, D., Tschardtke, T., 2007. Diversity of flower-  
12 visiting bees in cereal fields: effects of farming system, landscape composition  
13 and regional context. J. Appl. Ecol. 44, 41–49. doi:10.1111/j.1365-  
14 2664.2006.01259.x

15 Holzschuh, A., Steffan-Dewenter, I., Tschardtke, T., 2008. Agricultural landscapes with  
16 organic crops support higher pollinator diversity. Oikos 117, 354–361.  
17 doi:10.1111/j.2007.0030-1299.16303.x

18 Holzschuh, A., Steffan-Dewenter, I., Tschardtke, T., 2010. How do landscape composition  
19 and configuration, organic farming and fallow strips affect the diversity of bees,  
20 wasps and their parasitoids? J. Anim. Ecol. 79, 491–500. doi:10.1111/j.1365-  
21 2656.2009.01642.x

22 Kearns, C.A., Inouye, D.W., Waser, N.M., 1998. Endangered mutualisms: The conservation  
23 of plant-pollinator interactions. Annu. Rev. Ecol. Syst. 29, 83–112.  
24 doi:10.1146/annurev.ecolsys.29.1.83

25 Kennedy, C.M., Lonsdorf, E., Neel, M.C., Williams, N.M., Ricketts, T.H., Winfree, R.,

1 Bommarco, R., Brittain, C., Burley, A.L., Cariveau, D., Carvalheiro, L.G., Chacoff,  
2 N.P., Cunningham, S.A., Danforth, B.N., Dudenhoefter, J.-H., Elle, E., Gaines, H.R.,  
3 Garibaldi, L.A., Gratton, C., Holzschuh, A., Isaacs, R., Javorek, S.K., Jha, S., Klein,  
4 A.M., Krewenka, K., Mandelik, Y., Mayfield, M.M., Morandin, L., Neame, L.A., Otieno,  
5 M., Park, M., Potts, S.G., Rundlöf, M., Saez, A., Steffan-Dewenter, I., Taki, H., Viana,  
6 B.F., Westphal, C., Wilson, J.K., Greenleaf, S.S., Kremen, C., 2013. A global  
7 quantitative synthesis of local and landscape effects on wild bee pollinators in  
8 agroecosystems. *Ecol. Lett.* 16, 584–599. doi:10.1111/ele.12082

9 Klein, A.-M., Vaissiere, B.E., Cane, J.H., Steffan-Dewenter, I., Cunningham, S.A., Kremen, C.,  
10 Tschardtke, T., 2007. Importance of pollinators in changing landscapes for world  
11 crops. *Proc. R. Soc. B-Biol. Sci.* 274, 303–313. doi:10.1098/rspb.2006.3721

12 Kohler, F., Verhulst, J., van Klink, R., Kleijn, D., 2008. At what spatial scale do high-quality  
13 habitats enhance the diversity of forbs and pollinators in intensively farmed  
14 landscapes? *J. Appl. Ecol.* 45, 753–762. doi:10.1111/j.1365-2664.2007.01394.x

15 Kremen, C., Williams, N.M., Bugg, R.L., Fay, J.P., Thorp, R.W., 2004. The area requirements  
16 of an ecosystem service: crop pollination by native bee communities in California.  
17 *Ecol. Lett.* 7, 1109–1119. doi:10.1111/j.1461-0248.2004.00662.x

18 Kremen, C., Williams, N.M., Thorp, R.W., 2002. Crop pollination from native bees at risk  
19 from agricultural intensification. *Proc. Natl. Acad. Sci. U. S. A.* 99, 16812–16816.  
20 doi:10.1073/pnas.262413599

21 Kunin, W.E., 1997. Population biology and rarity: on the complexity of density-  
22 dependence in insect-plant interactions, in: Kunin, W.E., Gaston, K.J. (Eds.), *The*  
23 *Biology of Rarity: Causes and Consequences of Rare-Common Differences.*  
24 Chapman and Hall, London, pp. 150–173.

25 Nakagawa, S., Schielzeth, H., 2013. A general and simple method for obtaining  $R^2$  from

1 generalized linear mixed-effects models. *Methods Ecol. Evol.* 4, 133–142.  
2 doi:10.1111/j.2041-210x.2012.00261.x

3 Norton, L., Johnson, P., Joys, A., Stuart, R., Chamberlain, D., Feber, R., Firbank, L., Manley,  
4 W., Wolfe, M., Hart, B., Mathews, F., MacDonald, D., Fuller, R.J., 2009.  
5 Consequences of organic and non-organic farming practices for field, farm and  
6 landscape complexity. *Agric. Ecosyst. Environ.* 129, 221–227.  
7 doi:10.1016/j.agee.2008.09.002

8 Ockinger, E., Smith, H.G., 2007. Semi-natural grasslands as population sources for  
9 pollinating insects in agricultural landscapes. *J. Appl. Ecol.* 44, 50–59.  
10 doi:10.1111/j.1365-2664.2006.01250.x

11 Ollerton, J., Lack, A., 1998. Relationships between flowering phenology, plant size and  
12 reproductive success in *Lotus corniculatus* (Fabaceae). *Plant Ecol.* 139, 35–47.  
13 doi:10.1023/A:1009798320049

14 Ollerton, J., Winfree, R., Tarrant, S., 2011. How many flowering plants are pollinated by  
15 animals? *Oikos* 120, 321–326. doi:10.1111/j.1600-0706.2010.18644.x

16 Pellissier, V., Muratet, A., Verfaillie, F., Machon, N., 2012. Pollination success of *Lotus*  
17 *corniculatus* (L.) in an urban context. *Acta Oecologica-Int. J. Ecol.* 39, 94–100.  
18 doi:10.1016/j.actao.2012.01.008

19 Potts, S.G., Biesmeijer, J.C., Kremen, C., Neumann, P., Schweiger, O., Kunin, W.E., 2010.  
20 Global pollinator declines: trends, impacts and drivers. *Trends Ecol. Evol.* 25,  
21 345–353. doi:10.1016/j.tree.2010.01.007

22 Ricketts, T.H., Regetz, J., Steffan-Dewenter, I., Cunningham, S.A., Kremen, C., Bogdanski,  
23 A., Gemmill-Herren, B., Greenleaf, S.S., Klein, A.M., Mayfield, M.M., Morandin, L.A.,  
24 Ochieng, A., Viana, B.F., 2008. Landscape effects on crop pollination services: are  
25 there general patterns? *Ecol. Lett.* 11, 499–515. doi:10.1111/j.1461-

1           0248.2008.01157.x

2   Robinson, R.A., Sutherland, W.J., 2002. Post-war changes in arable farming and  
3           biodiversity in Great Britain. *J. Appl. Ecol.* 39, 157–176. doi:10.1046/j.1365-  
4           2664.2002.00695.x

5   Rundlöf, M., Bengtsson, J., Smith, H.G., 2008a. Local and landscape effects of organic  
6           farming on butterfly species richness and abundance. *J. Appl. Ecol.* 45, 813–820.  
7           doi:10.1111/j.1365-2664.2007.01448.x

8   Rundlöf, M., Nilsson, H., Smith, H.G., 2008b. Interacting effects of farming practice and  
9           landscape context on bumblebees. *Biol. Conserv.* 141, 417–426.  
10          doi:10.1016/j.biocon.2007.10.011

11   Rundlöf, M., Smith, H.G., 2006. The effect of organic farming on butterfly diversity  
12          depends on landscape context. *J. Appl. Ecol.* 43, 1121–1127. doi:10.1111/j.1365-  
13          2664.2006.01233.x

14   Scherber, C., Eisenhauer, N., Weisser, W.W., Schmid, B., Voigt, W., Fischer, M., Schulze, E.-  
15          D., Roscher, C., Weigelt, A., Allan, E., Bessler, H., Bonkowski, M., Buchmann, N.,  
16          Buscot, F., Clement, L.W., Ebeling, A., Engels, C., Halle, S., Kertscher, I., Klein, A.-M.,  
17          Koller, R., Koenig, S., Kowalski, E., Kummer, V., Kuu, A., Lange, M., Lauterbach, D.,  
18          Middelhoff, C., Migunova, V.D., Milcu, A., Mueller, R., Partsch, S., Petermann, J.S.,  
19          Renker, C., Rottstock, T., Sabais, A., Scheu, S., Schumacher, J., Temperton, V.M.,  
20          Tscharntke, T., 2010. Bottom-up effects of plant diversity on multitrophic  
21          interactions in a biodiversity experiment. *Nature* 468, 553–556.  
22          doi:10.1038/nature09492

23   Steffan-Dewenter, I., Potts, S.G., Packer, L., 2005. Pollinator diversity and crop pollination  
24          services are at risk. *Trends Ecol. Evol.* 20, 651–652.  
25          doi:10.1016/j.tree.2005.09.004

- 1 Steffan-Dewenter, I., Tscharntke, T., 1999. Effects of habitat isolation on pollinator  
2 communities and seed set. *Oecologia* 121, 432–440.  
3 doi:10.1007/s004420050949
- 4 Wilcock, C., Neiland, R., 2002. Pollination failure in plants: why it happens and when it  
5 matters. *Trends Plant Sci.* 7, 270–277. doi:10.1016/S1360-1385(02)02258-6
- 6 Williams, N.M., Kremen, C., 2007. Resource distributions among habitats determine  
7 solitary bee offspring production in a mosaic landscape. *Ecol. Appl.* 17, 910–921.  
8 doi:10.1890/06-0269
- 9 Winfree, R., Aguilar, R., Vazquez, D.P., LeBuhn, G., Aizen, M.A., 2009. A meta-analysis of  
10 bees' responses to anthropogenic disturbance. *Ecology* 90, 2068–2076.  
11 doi:10.1890/08-1245.1
- 12 Winfree, R., Griswold, T., Kremen, C., 2007. Effect of human disturbance on bee  
13 communities in a forested ecosystem. *Conserv. Biol.* 21, 213–223.  
14 doi:10.1111/j.1523-1739.2006.00574.x
- 15 Winfree, R., Williams, N.M., Gaines, H., Ascher, J.S., Kremen, C., 2008. Wild bee pollinators  
16 provide the majority of crop visitation across land-use gradients in New Jersey  
17 and Pennsylvania, USA. *J. Appl. Ecol.* 45, 793–802. doi:10.1111/j.1365-  
18 2664.2007.01418.x
- 19 Wolf, S., Moritz, R.F.A., 2008. Foraging distance in *Bombus terrestris* L. (Hymenoptera :  
20 Apidae). *Apidologie* 39, 419–427. doi:10.1051/apido:2008020

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22



## FIGURE CAPTIONS

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2

3 Figure 1: Relationship between *L. corniculatus* fruit set and landscape variables. Fruit set is  
4 the proportion of flowers setting fruit. Each point represents one pot (hence a maximum of 4  
5 points per site). The two landscape variables were the proportion of semi-natural habitat in a  
6 250-m radius buffer around a site (a) and the distance to the nearest patch of semi-natural  
7 habitat (b).

Table 1

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<b>Tables</b>
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Table 1: Mixed-effects modelling of the effects of local and landscape variables on fruit and seed set of *L. corniculatus*. Stars indicate significance levels: \* =  $0.01 < P < 0.05$ ; \*\* =  $0.001 < P < 0.01$ ; \*\*\* =  $P < 0.001$ .  $R^2_{\text{GLMM}(m)}$  and  $R^2_{\text{GLMM}(c)}$  are the marginal and conditional  $R$ -squared, i.e. the variance explained by fixed factors alone or by fixed and random factors, respectively in generalized linear mixed-effects models (Nakagawa and Schielzeth, 2013).

	Fruit set		Seed set	
<i>Fixed factors</i>	Estimate	[95% CI]	Estimate	[95% CI]
Intercept	-7.99	[-11.1;-4.86]	1.66	[1.08;2.25]
Farming system (conventional)	0.20	[-0.70;1.11]	0.24	[-0.079;0.55]
Local habitat type (set-aside)	<b>1.35**</b>	<b>[0.42;2.28]</b>	0.23	[-0.07;0.52]
Distance to semi-natural habitats	<b>-0.0039*</b>	[-7.3 10 <sup>-2</sup> ;-4.6 10 <sup>-4</sup> ]	7.2 10 <sup>-4</sup>	[-1.7 10 <sup>-3</sup> ;2.5 10 <sup>-4</sup> ]
Proportion semi-natural habitats	<b>-6.25**</b>	<b>[-9.9;-2.6]</b>	-0.84	[-2.0;0.31]
Variety (Leo)	-0.25	[-1.3;0.76]	-0.11	[-0.49;0.27]
Number of flowers	0.070	[0.023;0.12]	2.2 10 <sup>-3</sup>	[-2.7 10 <sup>-3</sup> ;7.110 <sup>-3</sup> ]
Number of varieties	<b>4.51***</b>	<b>[2.6;6.4]</b>	0.057	[-0.24;0.35]
Nb. flowers × Nb. varieties	<b>-0.039**</b>	<b>[-0.069;-0.009]</b>		
<i>Random factors</i>	Variance components		Variance components	
Farm pair	0.726		<10 <sup>-4</sup>	
Site	<10 <sup>-4</sup>		<10 <sup>-4</sup>	
Pot	1.202		0.0693	
Inflorescence	-		0.203	
$R^2_{\text{GLMM}(m)}$	38.95%		11.24%	
$R^2_{\text{GLMM}(c)}$	50.76%		66.71%	

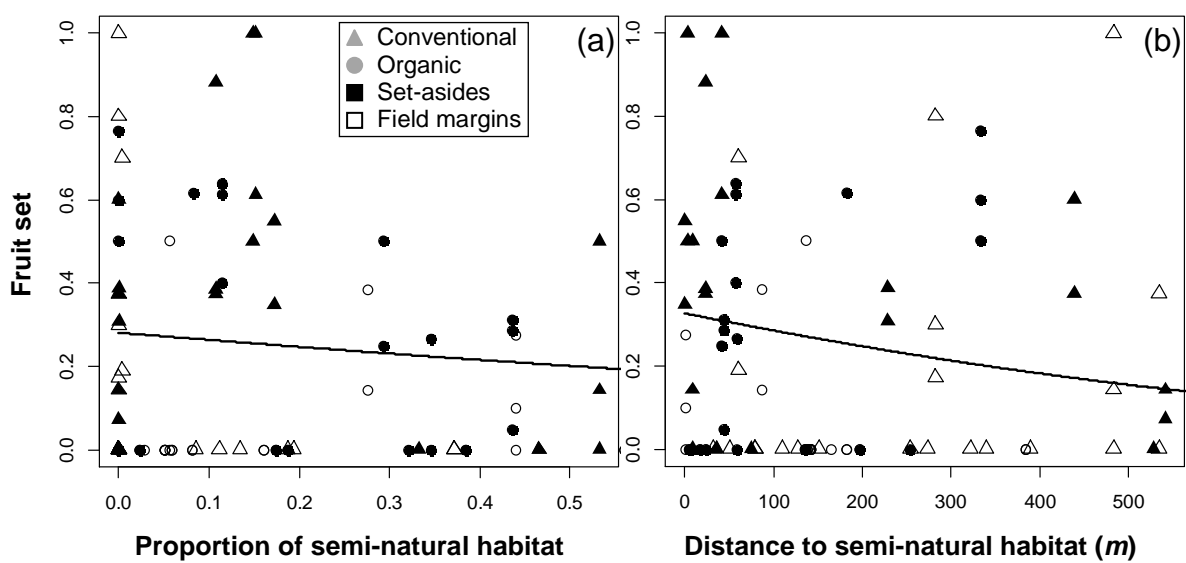


Figure 1